

Rahman M., Asaeda T., Abeynayaka H.D.L., Fukahori K. (2023), The long-term effect of light intensities in different temperatures can play a crucial role in suppressing *Microcystis aeruginosa*, pp. 17-31. In Gastescu, P., Bretcan, P. (edit., 2023), Water resources and wetlands, 6th International Hybrid Conference Water resources and wetlands, 13-17 September 2023, Tulcea (Romania). pp. 287

Available online at <http://www.limnology.ro/wrw2023/proceedings.html>

Open access under CC BY-NC-ND license

6th International Hybrid Conference Water resources and wetlands, 13-17 September 2023, Tulcea (Romania)



THE LONG-TERM EFFECT OF LIGHT INTENSITIES IN DIFFERENT TEMPERATURES CAN PLAY A CRUCIAL ROLE IN SUPPRESSING *MICROCYSTIS AERUGINOSA*

Mizanur RAHMAN^{1*}, Takashi ASAEDA^{1,2,3*}, Helayaye Damitha Lakmali ABEYNAYAKA¹, Kiyotaka FUKAHORI¹

¹ Saitama University, Saitama 338-8570, Japan

² Hydro Technology Institute, Shimo-meguro, Tokyo Japan

³ Research and Development Center, Nippon Koei, Tsukuba, Japan

Email IDs: M.R. (masudbiochem2012@gmail.com); T.A. (asaedat@gmail.com); H.D.L.A.

(hdlakmali@yahoo.com); K.F (fukahori@mail.saitama-u.ac.jp)

*Correspondence and requests for materials should be addressed to T.A. (asaedat@gmail.com) and M.R. (masudbiochem2012@gmail.com)

Abstract. *Microcystis aeruginosa* toxicity is an increasing problem for human and animal health. Therefore, this study is designed to evaluate the long-term effect of light intensities in different temperatures to eradicate *M. aeruginosa*. *M. aeruginosa* was acclimated at 30°C and 20°C in two incubators for 24 days. After 24 days of acclimatization, the first samples were collected (known as zero days), and the second samples were collected seven days later for the five light intensities, 10, 30, 50, 200, and 600 $\mu\text{mol m}^{-2} \text{s}^{-1}$. The samples were collected to analyze optical density (OD₇₃₀), H₂O₂ concentration, protein content, chlorophyll-a (Chl-a) content, and the antioxidative enzyme activities of catalase (CAT) at 30°C and 20°C. To homogenize *M. aeruginosa* cells, five to eight 3mm beads (BMS Inc., Japan) were used with the appropriate buffer and vigorously shaken. The zero days collection sample is identical because preliminary light intensities were changed during the collection time in each temperature. In the seven-day exposure compared to zero-day, H₂O₂ concentration and protein content showed a parallel variational increasing trend until 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$ of light intensity, then decreasing afterward at each temperature and were higher at 30°C than 20°C. Chl-a drastically decreased from 50 $\mu\text{mol m}^{-2} \text{s}^{-1}$ to 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$ and continued thereafter in the seven days of treatment. Higher H₂O₂ concentration decreases Chl-a content from zero days to seven days in both temperatures. CAT activity is proportionate to H₂O₂ and H₂O₂/protein content in both temperatures in various light intensities. Photoinhibition occurs at 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$ after seven days of exposure at 30°C and 20°C as algal biomass (H₂O₂/protein) decreases. Algal biomass is higher at 30°C than 20°C indicating 30°C is the more favorable temperature for them to grow. Therefore, H₂O₂ can be used as a possible biomarker. These results indicate that high light intensity (~ 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$), becomes oxidative stresses in both temperatures that can reduce algal biomass. These findings can be used as a nonchemical approach to eliminating *M. aeruginosa*.

Keywords: *Microcystis aeruginosa*, Hydrogen peroxide, temperatures, protein content, light intensities

1 INTRODUCTION

Many countries experience the bloom of cyanobacterium *Microcystis aeruginosa* in eutrophic lakes and reservoirs, including Japan (Dokulil and Teubner 2000; Ma and Wang 2021; Ohtsuka 2000; Ozawa et al. 2003). *M. aeruginosa* has been reported as toxic to mammals (Carmichael 1981; Gorham 1964a; Schwimmer and Schwimmer 1968). Some of the strains of *M. aeruginosa* produce hepatotoxins called microcystins (Otsuka et al., 1999 and 2000; Rantala et al., 2004; Tillett et al., 2001; Yoshida et al., 2008), which cause hepatocellular carcinoma (MacKintosh et al., 1990; Yoshizawa et al 1990). Various factors, such as temperature and light, affect the distribution and abundance of cyanobacteria (Cao et al. 2006, 2008; Joehnk et al. 2008; Asaeda et al. 2022a).

Compared to other algae species, *M. aeruginosa* can proliferate at high temperatures, making cyanobacteria prevalent in summer blooms (Drake et al., 2010). The preferable growth temperature for *Microcystis spp* is from 24°C to 34°C (Ganf, 1974). It is suggested that changing water temperature can effectively control cyanobacterial blooms (Chu et al. 2007). However, there has been little documentation of the species-specific effect of temperature change on their growth.

Cyanobacteria are directly dependent upon light. Stress may be caused by moderate light changes (Welkie et al. 2019). Cyanobacteria rely almost exclusively on light for energy, so responding to various light intensities is essential (Asaeda et al., 2022a). A lot can be learned about photosynthesis and biotechnological applications of cyanobacteria based on how they respond to various light intensities (Flombaum 2013; Oliver 2016; Leblanc Renaud et al. 2011). The light intensity changes within a day also affect the habitat preference of cyanobacteria (Saha et al. 2016). Through an elaborate electron transport pathway, cyanobacteria carry out photosynthesis and respiration with the help of solar energy (Lea-Smith et al., 2006). Excess solar energy produces reactive oxygen species (ROS), including superoxide radicals (Latifi et al., 2009; Raja et al., 2017), and damages cellular components, such as the D1 protein. This protein is responsible for recovering the damaged photosynthesis apparatus (Gill and Tuteja 2010; Nishiyama, Y., Murata, 2014; Weerakoon et al., 2018). Antioxidant activities detoxify superoxide radicals into the water through superoxide dismutase (Asada et al., 1999; Ma and Gao 2010; Rastogi et al., 2010; Sharma et al., 2012) and photoinhibition occurs. Thus, excessive solar radiation inhibits the proliferation of cyanobacteria. In the ecology of cyanobacteria in water, even though solar radiation intensity is highest at the surface, photoinhibition plays an important role in reducing cyanobacterial biomass (Harriss and Smith 2013; Salonen et al., 1999).

The eradication or prevention of cyanobacteria growth has been the subject of numerous studies worldwide (Paerl et al. 2011; Rodríguez-Molares et al. 2014; Rajasekhar et al. 2012). The use of chemical methods to eliminate cyanobacteria is discouraged due to their secondary harm to ecosystems (Grandgirard et al. 2002; Lake et al. 2003; Jančula et al. 2011). Consequently, non-chemical approaches can be used, but they require more understanding of the relationship between cyanobacteria and the environment, their interaction with other species, and their responses to environmental stresses (Asaeda et al. 2022a).

Living organisms and their ecosystems play a crucial role in preventing or repairing the damage caused by climate change. Stress triggers the production of endogenous reactive oxygen species (ROS) (Caverzan et al., 2012; Sharma et al., 2012). The overproduction of ROS can be controlled by antioxidant activities. When the antioxidants' capacity to scavenge ROS is exceeded by oxidative stress, ROS imbalance will occur (Dumont et al., 2019; Ugya et al., 2020). Several different environmental stressors, such as various light intensities or temperatures (Leblanc Renaud et al. 2011; Wu et al. 2008), generate H₂O₂, superoxide radical (O₂⁻), as well as the hydroxyl radical (OH⁻) (Sharma et al. 2012; Asada et al. 2006; Satterfield et al. 1955; Zhou et al. 2006). Most of the ROS will be converted to H₂O₂. Some of them will be extremely dangerous, although antioxidant activities have also increased to reduce the risk of damage to the cyanobacteria cells (Asaeda et al. 2018, 2019, 2020, 2022a; Barnuevo and Asaeda, 2018; Rahman et al. 2023a, 2023b).

While light intensity and temperature have been studied independently by numerous researchers (Wu et al., 2008; Leblanc Renaud et al. 2011; Utkilen et al. 1992; Asaeda et al. 2020, 2021), little information is available about their combined effects on cyanobacterial growth. Therefore, the main objective of this study is to investigate the long-term combined effect of temperatures and different light intensities for cyanobacterial bloom treatment as a non-chemical approach.

2 MATERIALS AND METHOD

2.1 *M. aeruginosa* cell culture

M. aeruginosa (NIES-111) is a bloom-forming blue-green algae that usually form colonies obtained from the National Institute for Environmental Studies (NIES), Japan. The Autoclaved 100% BG 11 (Rippka, et al. 1979) medium was used for the *M. aeruginosa* culture throughout the experiment. The strain was cultured at 30°C (± 0.3°C) with white fluorescent light, containing a flux of 10–20 μmol m⁻² s⁻¹ in a 12 h:12 h light: dark cycle inside an incubator (Cool Incubator KMH-259 from AS ONE Corporation). Each culture was manually shaken twice a day during the light phase in the incubation period. *M. aeruginosa* was subcultured (Abeynayaka et al. 2017, 2018) until a sufficient amount was achieved.

2.2 Experimental procedure

A sufficient amount of *M. aeruginosa* was transferred into two incubators at 30°C ($\pm 0.3^\circ\text{C}$) and 20°C ($\pm 0.3^\circ\text{C}$) at 10–20 $\mu\text{mol m}^{-2} \text{s}^{-1}$ photosynthetically active radiation (PAR) for 24 days to acclimatize. To examine the long-term effect of light exposure in *M. aeruginosa* at 30°C and 20°C, five different light intensities—namely, 10, 30, 50, 200, 100 and 600 $\mu\text{mol m}^{-2} \text{s}^{-1}$, were introduced by white fluorescent light (Toshiba, Japan). An automatic timer (REVEX PT7, Saitama, Japan) was set to maintain 12 hours of light and 12 hours of darkness. The light intensities were measured using a quantum sensor (Apogee, MQ-200, United States) and adjusted uniformly in the media. Cyanobacteria culture flasks were gently shaken three times daily to ensure homogeneous exposure to the light environment. Samples were taken twice for analysis. The first samples were taken at 12:00 on the 24 days of acclimatization, considered zero days. Following the first sampling, a second sampling was conducted at 12:00 with a 7-day interval. Collected samples were allocated in 1.5 mL tubes, taking 1 mL of *M. aeruginosa* cells, and centrifuged at 10,000 rotations per minute (rpm) for 10 minutes at 4°C. The supernatant was removed and kept at -80°C for bioassays described below. *M. aeruginosa* cell wall cannot break down by using the buffer (phosphate buffer (pH7; 0.1M), NaOH solution for measuring H₂O₂ and protein respectively). Therefore, five to eight 3mm beads (BMS Inc., Japan) were added with buffer and vigorously shaken. The mixture was centrifuged at 10,000 rpm for 10 min at 4 °C and the supernatant was used as an extract for further analysis. Each culture group was replicated three times to conduct the experiment.

2.3 Measurement of total soluble protein concentration

Minor modifications were made to (Abeynayaka et al., 2017)'s method to estimate total soluble protein concentrations. Cyanobacterial cells were unfrozen, and distilled water was used to wash the pellet once. The extraction procedure was performed using 1mL of 0.5 M NaOH solution using beads. The extraction was centrifuged at 4°C for 15 minutes at 10,000 rpm to pick up the supernatant. A Coomassie Bradford protein assay was used to measure protein content. After incubation at room temperature ($25 \pm 2^\circ\text{C}$) for 10 min, the absorbance was recorded at 595 nm using a UV-Vis spectrophotometer (UVmini-1240, Shimadzu, Japan). Protein levels were determined by a series of known albumin concentrations.

2.4 Chlorophyll-a concentration analysis

Using (ISO 10260:1992; Gregor and Marsalek, 2003)'s method, Chl-a concentrations were measured in cyanobacteria samples. The pellet was washed with distilled water once after each cyanobacterial cell was unfrozen, then extracted in 1.5 mL of 90% ethanol. After vigorous shaking, the mixture was kept at room temperature ($25 \pm 2^\circ\text{C}$) overnight in darkness. Then, each sample was centrifuged at 12,000 rpm, and the supernatant was measured with a UV-Vis spectrophotometer (UVmini-1240, Shimadzu, Japan) at the absorption wavelengths of 665nm and 649nm. The Chl-a content was calculated by using the following equation:

$$\text{Chl-a} = (13.95 * A_{665} - 6.88 * A_{649}) * 1.5$$

2.5 Identification of cell growth by measuring OD₇₃₀

OD₇₃₀ measurements were performed on each flask in order to determine the cyanobacteria growth rate. With a UV-Vis spectrophotometer (UVmini-1240, Shimadzu, Japan), OD₇₃₀ was measured at a wavelength of 730 nm using a previously proposed methodology (Axler et al., 1994; Taras 1971)

2.6 H₂O₂ concentration measurement assay

A modified version of xylenol orange (eFOX) method was used to measure the H₂O₂ content of *M. aeruginosa* following the methods of (Queval et al., 2008; Cheeseman 2006). The homogenized cells were centrifuged at 10,000 rpm for 10 minutes at 4°C to extract the cellular H₂O₂. The supernatant (100 μL) was added to 1 mL of the assay solution containing 0.250 mM ferrous ammonium sulphate, 0.1 mM sorbitol, 0.1 mM xylenol orange, 1% ethanol, and 25 mM H₂SO₄, which had been deoxygenated with gaseous nitrogen to prevent artifact production in hydrogen peroxide during the reaction. After 15 minutes of reaction, the absorbance at 560 nm was measured by spectrophotometry. H₂O₂ content was calculated by a standard curve using a series of diluted solutions of commercial, high-grade 9.8 M H₂O₂ (Dautania et al. 2014; Asaeda et al. 2022a).

2.7 CAT assay

CAT activity was measured using the method described by (Aebi, 1984). A total of 1 mL of unfrozen cell pellets were homogenized in 1 mL potassium phosphate buffer (0.05 M, pH 7.0) containing 0.1 mM EDTA. The supernatant was collected as an enzymatic extract after centrifugation (12,000 rpm at 4°C for 10 min). The CAT activity was measured by reacting 15 μL of 0.75M H_2O_2 , 920 μL of potassium phosphate buffer, and 65 μL of extract supernatant. Optical absorption was measured at 240 nm using UV mini-1240. The measurements were recorded every 10 seconds for 3 minutes, and the CAT activity was computed using a 39.4 mM/cm extinction coefficient.

2.8 Statistical Analysis

A one-way analysis of variance (ANOVA), followed by the Tukey test, was performed to check the statistical significance of variations among the means of sample groups. Significant differences between experimental groups of *M. aeruginosa* were evaluated using an independent sample t-test, assuming equality of variance. Pearson's correlation analysis was used to assess correlations among parameters. Statistical analyses were performed using IBM SPSS Statistics for MacBook, Version 28.0 (IBM Corporation, USA).

3 RESULTS

Due to preliminary light intensities changing during collection, the zero days collection in each temperature did not show too much fluctuation. Figure 1 shows the differences in OD_{730} between 30°C and 20°C regarding different light intensities. OD_{730} values significantly increased from 0 to 7 days at 20°C ($p < 0.001$). In the 7 days at 30°C, OD_{730} decreased with increasing PAR. In the meantime, OD_{730} increased till 30 PAR at 20°C and took nearly the same value thereafter. A positive correlation was found between OD_{730} and different light intensities in 7 days between 30°C and 20° ($r = 0.417$). Protein concentration substantially increased from 0 days to 7 days at 20 °C ($p < 0.001$), whereas at 30°C, it slightly increased until 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$, declined with 600 $\mu\text{mol m}^{-2} \text{s}^{-1}$ (Figure 2). With respect to light intensity, after 7 days of light exposure, the protein concentration increased until 30 $\mu\text{mol m}^{-2} \text{s}^{-1}$, then constantly decreased with increasing light intensities, similarly in 30 °C and 20°C experiments. After 7 days of treatment, H_2O_2 concentration and H_2O_2 /protein or algal biomass significantly increased in both temperatures ($p < 0.001$ for each temperature) (Figure 3 and Figure 4). At 7 days, H_2O_2 and H_2O_2 /protein concentration increased to 200 PAR at each temperature, and then it decreased with higher PAR. A significant positive correlation was observed from 30°C to 20°C between PAR with H_2O_2 ($r = 0.859$ $p < 0.001$) and PAR with H_2O_2 /protein ($r = 0.914$ $p < 0.001$). In both temperatures, Chl-a increased until 50 PAR and continued to decrease with higher PAR (Figure 5). The Chl-a concentration is shown as a function of H_2O_2 of *M. aeruginosa* under different light intensities in Figure 6 in both temperatures. With 20°C, Chl-a concentration increased with less than 0.5 mmol/L of H_2O_2 . However, with a higher concentration of H_2O_2 , it decreased. With 30°C, Chl-a concentration slightly increased until 1 mmol/l of H_2O_2 , then decreased with a higher concentration of H_2O_2 . CAT activity significantly increased from 0 days to 7 days at 30°C ($p < 0.01$) and 20°C ($p < 0.01$) to decrease oxidative stress. CAT activity increased to 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$ at each temperature. CAT activity and temperatures showed a positive correlation (30°C and 20°C, $r = 0.415$).

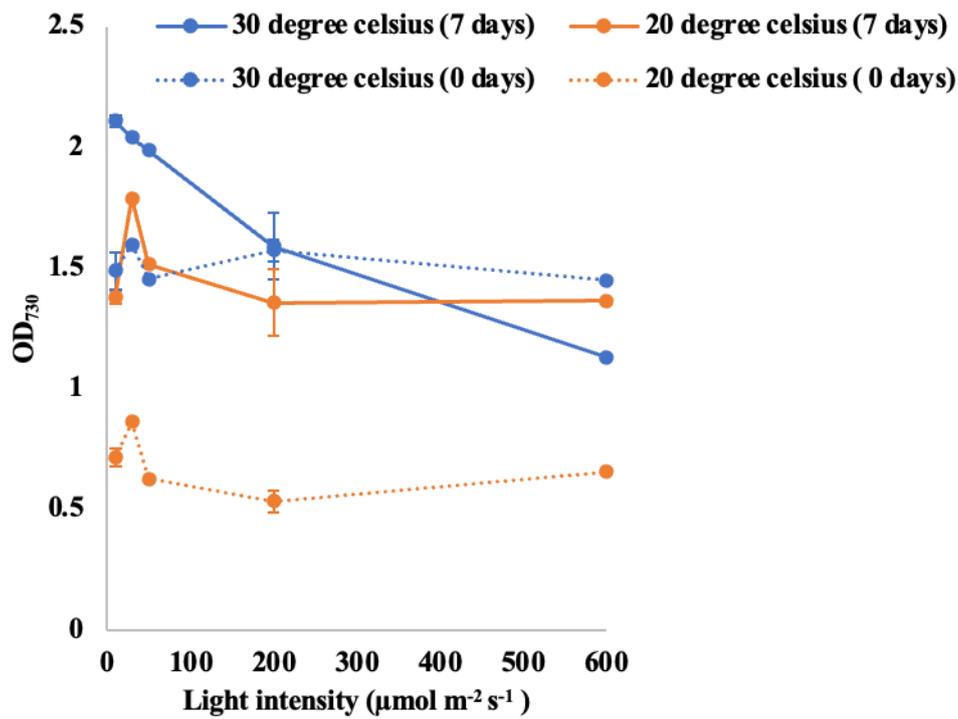


Figure 1. *M. aeruginosa* OD₇₃₀ at 0 and 7 days at 30°C and 20°C. Solid lines indicate seven-day treatment, whereas dashed lines indicate 0 days. The error bars indicate standard deviations.

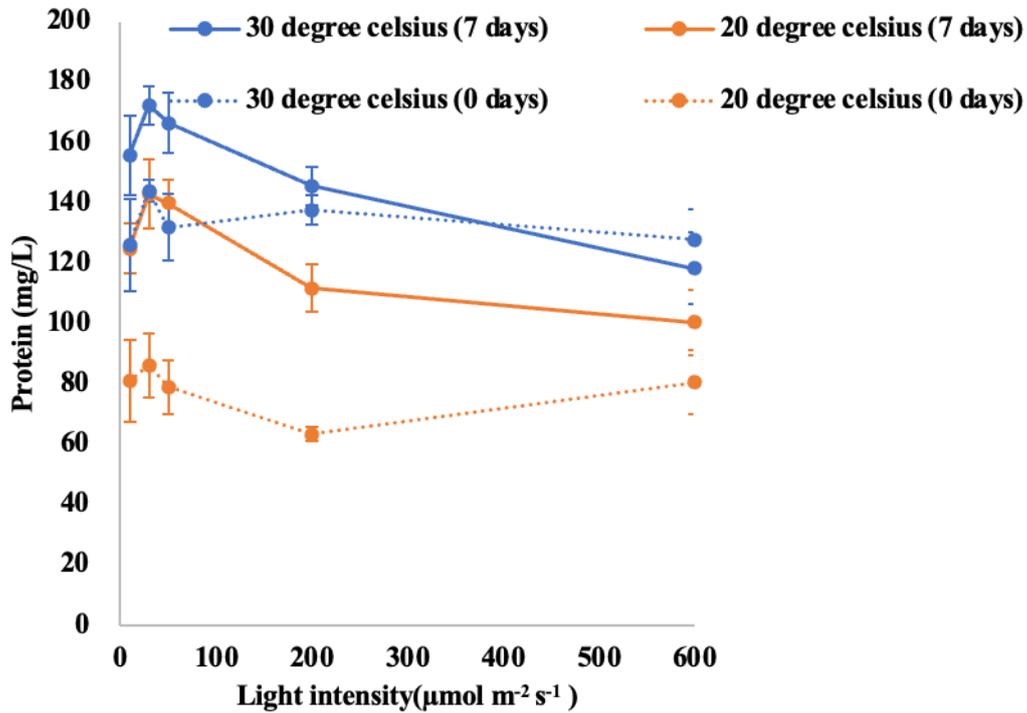


Figure 2. Changes in protein concentration concerning light intensity at 0 and 7 days of observing *M. aeruginosa*. Solid lines indicate seven-day treatment, whereas dashed lines indicate 0 days. The error bars indicate standard deviations.

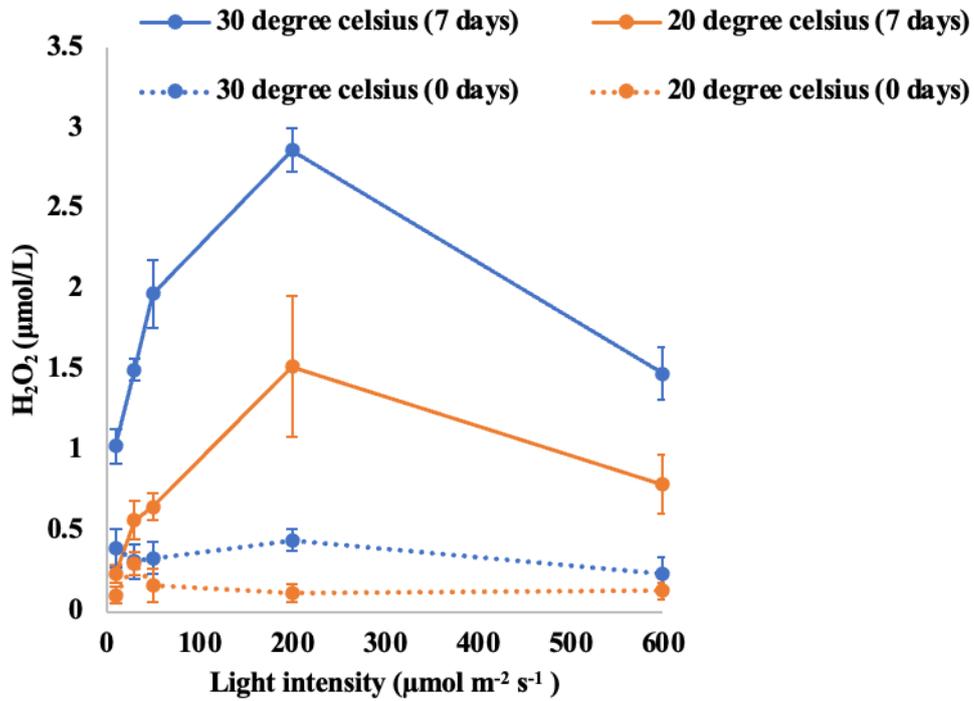


Figure 3. Changes in H₂O₂ concentration concerning light intensity in zero and seven days of observing *M. aeruginosa*. Solid lines indicate seven-day treatment, whereas dashed lines indicate 0 days. The error bars indicate standard deviations.

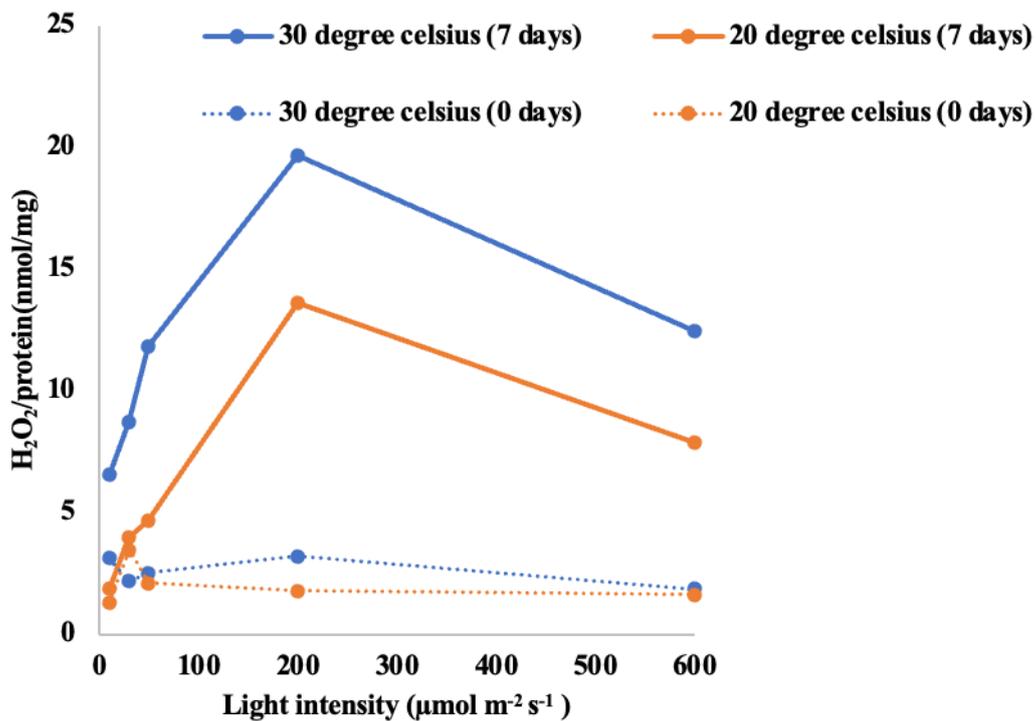


Figure 4. Changes in H₂O₂/protein concentration concerning light intensity at zero and seven days of observing *M. aeruginosa*. Solid lines indicate seven-day treatment, whereas dashed lines indicate 0 days. The error bars indicate standard deviations.

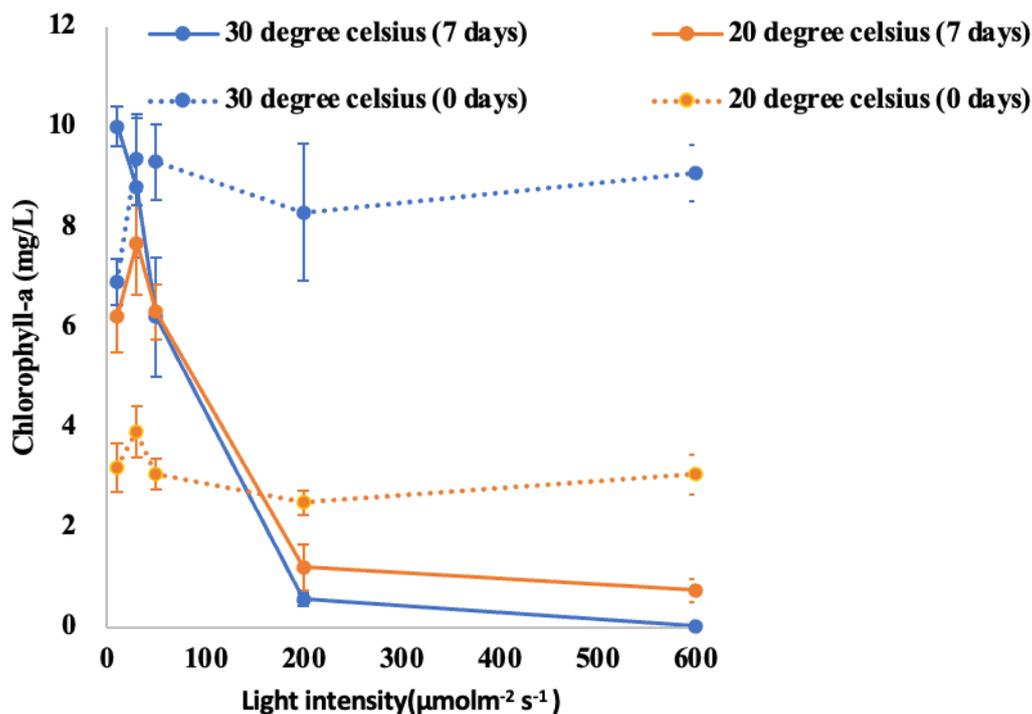


Figure 5. Changes in chlorophyll-a content concerning light intensity in zero and seven days of *M. aeruginosa* observation. Solid lines indicate seven-day treatment, whereas dashed lines indicate 0 days. The error bars indicate standard deviations.

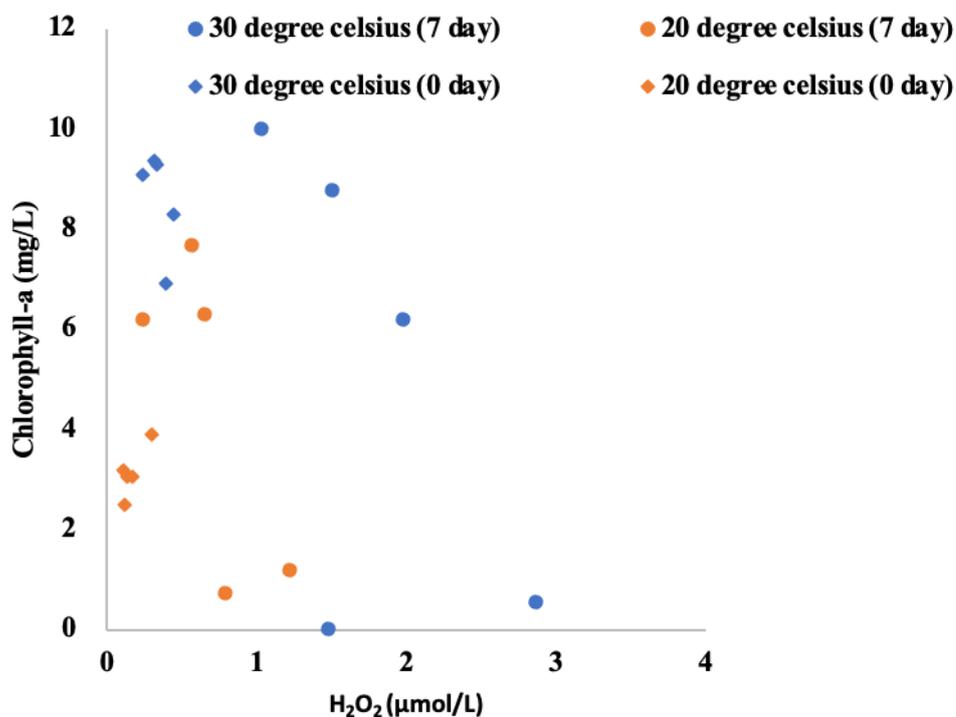


Figure 6. Changes in chlorophyll-a content concerning light intensity at zero and seven days of *M. aeruginosa* observation. Different symbols are used to separate 0 and 7 days of exposure. The error bars indicate standard deviations.

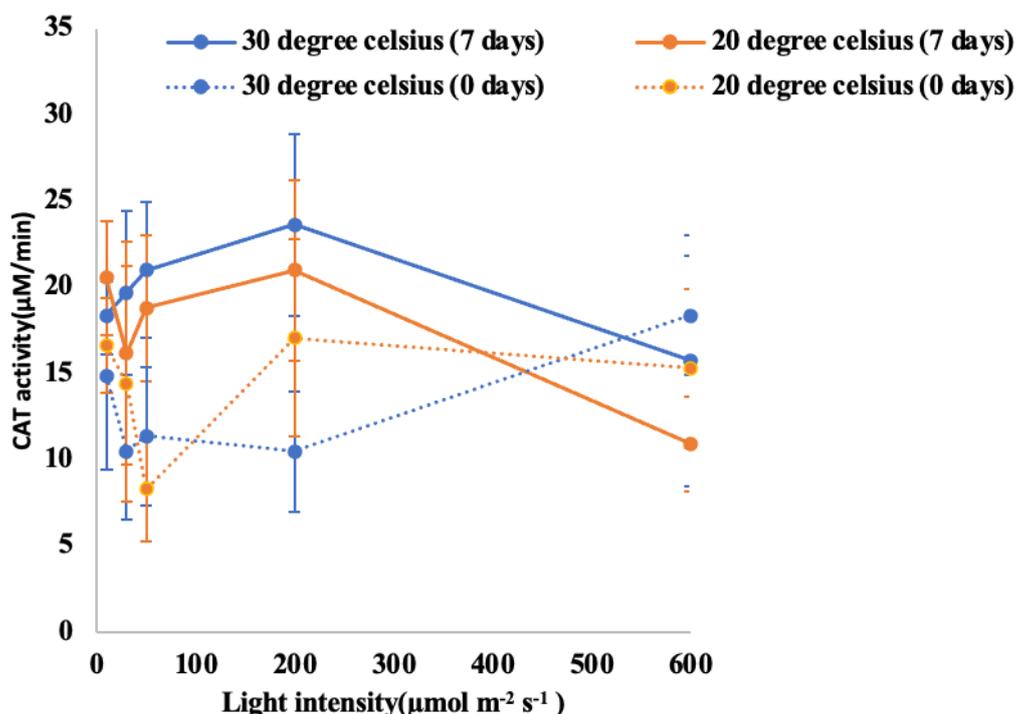


Figure 7. Changes in CAT activity concerning light intensity at zero and seven days of *M. aeruginosa* observation. Solid lines indicate seven-day treatment, whereas dashed lines indicate zero days. The error bars indicate standard deviations.

4 DISCUSSION

The accumulation of ROS is reported to increase in parallel with increased abiotic stress (Vestervik et al., 2012; Preece et al., 2017; Pham et al., 2018). In the present experiment, two abiotic stresses were applied, temperatures (30°C and 20°C) and low to high light intensities (10 to 600 $\mu\text{mol m}^{-2} \text{s}^{-1}$). Though H_2O_2 is produced under normal environmental conditions, its production is accelerated under high light intensity. In natural water, cyanobacteria are exposed to various abiotic stresses that enhance oxidative stress, producing H_2O_2 , which may deteriorate cyanobacterial biomass by producing hydroxyl radicals. It is not necessarily cumulative that H_2O_2 is produced due to different abiotic stresses (Mittler et al., 2006; Saints et al., 2010). However, the H_2O_2 concentration was enhanced with increasing PAR intensity at a certain stage and decreasing at lower temperatures (Figure 2).

A significant positive correlation was found between OD_{730} and protein content after seven days of exposure at 30°C ($r=0.853$, $p<0.001$) and 20°C ($r=0.696$, $p<0.01$). This relationship indicates that the protein content is a proper biomass reference. It is reported that one-third to one-half of the cell biomass is composed of protein (López et al., 2010). Therefore, $\text{H}_2\text{O}_2/\text{protein}$ is an appropriate indicator of H_2O_2 content in a cell.

4.1 Effect of light intensities and temperatures

M. aeruginosa exhibited a fluctuating response in different PAR. *M. aeruginosa* reached maximum H_2O_2 concentration at 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$ at both temperatures and decreased with higher light intensities. Both parameters' light intensities and temperatures were stressful mainly due to photosystem-produced ROS, and it will be converted to H_2O_2 . The present results indicated that photoinhibition occurs even with 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$; however, in the field observation, photoinhibition occurred when the light intensity exceeded 1000 $\mu\text{mol m}^{-2} \text{s}^{-1}$ (Whitelam and cold 1983; Machová et al., 2008; Harel et al., 2004). Although surface colony receives high intensity of solar adiation, it declined relatively quickly inside water, particularly in the water rich in cyanobacterial biomass. Thus, the cyanobacteria colony does not receive high solar radiation directly. In addition, they avoid the highest solar radiation in a day, by migrating to deeper zones (Molle et al., 2005, Athukorala et al., 2010; Cañedo-Argüelles et al., 2013; McLellan et al., 2017), likely to avoid the high solar radiation and the oxidative stress before the recovery of homeostasis via the increasing antioxidant activities (Athukorala et al., 2010).

The growth of *M. aeruginosa* is significantly decreased by temperatures between 30°C and 20°C ($r=0.752$, $p<0.01$), similar to other studies (Chu et al. 2007; Imai et al. 2009; You et al. 2018). At the same

time, temperatures significantly increased both H_2O_2 and H_2O_2 /protein concentration with each light intensity (30 to 20, $p < 0.01$) until 200 PAR. In contrast, a positive correlation between CAT and H_2O_2 with temperatures (30°C and 20°C, $r = 0.415$) was also found. These results indicate that *M. aeruginosa* maintains higher metabolism, whether it is regarding oxidative stress or antioxidant activities, with higher temperature, and it brings about a high growth rate with high temperature.

4.2 Effect of total protein content

The total protein content decreased with increasing light intensities, particularly after 200 PAR, and chlorophyll-a extremely decreased even after 30 $\mu\text{mol m}^{-2} \text{s}^{-1}$ both at 30°C and 20°C. At the same time, H_2O_2 /protein decreased from 200 PAR onwards (Figure 2). The reduction of chlorophyll-a concentration is due to the breakdown of photosynthesis components, such as photosystems I and II (Kura-Hotta et al., 1987), phycobilisomes (Raps et al., 1985; Walsh et al., 1997), and it decreases photosynthetic production. Phycobiliprotein synthesis is reduced in *M. aeruginosa* when light intensities exceed 200 $\text{mol m}^{-2} \text{s}^{-1}$ to protect against the absorption of excess light energy and protein degradation by proteases. (De Oliveira et al., 2014; Pojidaeva et al., 2004; Rosales-Loaiza et al., 2008). This reveals that *M. aeruginosa* has a defense mechanism to prevent cell damage from light and temperatures. The negative trend of total proteins with increasing light may also have influenced microcystin content (Kaebnick et al., 2000).

OD_{730} of *M. aeruginosa* in temperature also showed a fluctuation at 20°C; meanwhile, it started decreasing from 10 $\mu\text{mol m}^{-2} \text{s}^{-1}$ to 600 $\mu\text{mol m}^{-2} \text{s}^{-1}$ at 30°C in the seven days, respectively. In addition, it showed that *M. aeruginosa* was capable of surviving and proliferating continuously, despite its relatively weak growth rate and strict preference for 200 $\text{mol m}^{-2} \text{s}^{-1}$, and experienced high stress under higher light intensities. However, the observed increases in protein content with increasing light stress suggest that there was a deviated stress response for *M. aeruginosa*. The increased protein content can be associated with the upregulation of stress-related protein, as the oxidative stress was enforced due to the elevated H_2O_2 content with increasing light intensity (Babele et al., 2019). However, further research focused on the upregulation of stress proteins is necessary to confirm this phenomenon.

4.3 Effect of antioxidant activity

By triggering the antioxidant defense system in cyanobacteria, the oxidative stress response mechanism protects them from adverse environmental conditions (Liu et al., 2017). Oxidative stress and antioxidative enzymes are disturbed by abiotic stress (De Silva and Asaeda, 2017; Rastogi et al., 2010). In the present study, high light intensities (200 to 600 $\mu\text{mol m}^{-2} \text{s}^{-1}$) led to a reduction in H_2O_2 concentration, and low light intensities (10 to 200 $\mu\text{mol m}^{-2} \text{s}^{-1}$) augmented it. *M. aeruginosa* showed increasing trends with CAT compared to zero to 7 days in each temperature. However, when testing the regression relationships between H_2O_2 /protein and antioxidant enzymes, CAT activity was found to have a strong relationship at both temperatures. This confirms that CAT played a stronger role in antioxidant activity. The antioxidant balance also declined under higher stress. The CAT/protein activity with H_2O_2 /protein concentration of *M. aeruginosa* under different light intensities is proportionate in both temperatures. CAT/protein values increased with the increment of H_2O_2 /protein (Figure not shown).

4.4 Potentiality of using H_2O_2 as a bio-marker

Many studies have been conducted on the lethal concentration of H_2O_2 in cyanobacteria, mainly via incubation in laboratories, with the endorsement of different concentrations of H_2O_2 (Derakhshan et al., 2018; Shekoohiyan et al., 2013; Drábková et al., 2007a; Bouchard et al., 2011 Leunert et al., 2014; Foo et al., 2020). Laboratory incubations under different H_2O_2 concentrations provided the lethal H_2O_2 dosage for cyanobacteria in the past, which implies that cyanobacterial biomass is degraded with higher H_2O_2 concentrations. In the present study, because chlorophyll-a can be expressed as protein, we used protein content instead of chlorophyll-a content as an indicator of biomass (Marsac 1977).

H_2O_2 is generated in the photolysis process of dissolved organic carbon in water by UV (Cooper and Lean 1989; Scully et al., 1996). A water body's nutrient content increases its production rate of H_2O_2 due to its widespread distribution (Herrmann 1996). Thus, naturally generated H_2O_2 may be able to control cyanobacterial biomass in water and contribute to the decline of cyanobacteria. The protein content in water declined with an increasing H_2O_2 /protein concentration until 200 PAR (Figure 4). A higher protein level was not observed with higher H_2O_2 concentration levels in the present study (Figure 2). CAT activities are directly proportional to H_2O_2 and H_2O_2 /protein concentration with the increasing PAR at both temperatures and then decrease afterward. Therefore, the growth of cyanobacteria is suppressed by the generation of higher H_2O_2 levels.

4.5 Introducing an environmentally favourable approach to algal treatment

As a non-chemical method to control *M. aeruginosa*, high light exposure levels and temperature fluctuation is highly appreciated compared with chemical methods. However, the hypothesis of the method, such that low-light exposure suppresses the growth of cyanobacteria, is opposite to the present results. Supposedly, artificial water mixing in lakes and reservoirs (Visser et al., 2016) is not necessarily efficiently operated. In the case of introducing the system, *M. aeruginosa*'s both low- and high-light vulnerability can be carefully considered. The most efficient destratification process must be designed from light intensity and temperature distribution points beforehand (Asaeda and Imberger, 1993; Imteaz and Asaeda, 2000). In order to control *M. aeruginosa*, there is a possibility to introduce other types of practical methods, such as exposing deep water to high lights without raising the temperature. Methods should be developed to illuminate the water column beyond the tolerable light intensity level for *M. aeruginosa*.

5 CONCLUSION

M. aeruginosa was well-suited to illumination ranges of 30 to 50 $\mu\text{molm}^{-2}\text{s}^{-1}$. Beyond the optimal light intensities, the growth of *M. aeruginosa* was reduced, particularly after 200 PAR. Both temperatures significantly affect H_2O_2 concentration and protein concentration. $\text{H}_2\text{O}_2/\text{protein}$ and $\text{CAT}/\text{protein}$ also increased until 200 $\mu\text{mol m}^{-2}\text{s}^{-1}$ and decreased to 600 $\mu\text{mol m}^{-2}\text{s}^{-1}$ at seven days in both temperatures. *M. aeruginosa* demonstrated a greater tolerance for high light intensities. It could be possible to develop control mechanisms based on high-light intensities, which might lead to lower growth rates, or to improve the existing methods that rely on low-light intensities. This could help to control cyanobacteria in water bodies effectively.

6 DATA AVAILABILITY STATEMENT

The authors declare that the data will be available for everyone upon reasonable request.

7 COMPETING INTERESTS

The authors declare that there is no conflict of interest.

8 AUTHOR CONTRIBUTIONS

M.R. made experiments and chemical analyses and completed the first draft. T.A. contributed to the design of the experiment and the revision of MS. H.D.L.A. contributed to experiments along with M.R. K.F. extensively revised the manuscript.

9 FUNDINGS

This work was financially supported by the Grant-in-Aid for Scientific Research (B) (19H02245), (C) (20K04714), and the Fund for the Promotion of Joint International Research (18KK0116) of Japan Society for the Promotion of Science (JSPS).

10 REFERENCES

- Aebi H. 1984. Catalase in vitro. In *Methods in enzymology*, **105**: 121-126. Academic press. [https://doi.org/10.1016/S0076-6879\(84\)05016-3](https://doi.org/10.1016/S0076-6879(84)05016-3).
- Abeynayaka H D L, Asaeda T, and Kaneko Y. 2017. Buoyancy limitation of filamentous cyanobacteria under prolonged pressure due to the gas vesicles collapse. *Environmental management*, **60**(2), 293-303. doi:10.1007/s00267-017-0875-7.
- Abeynayaka H D L, Asaeda T, and Rashid M H. 2018. Effects of elevated pressure on *Pseudanabaena galeata* Böcher in varying light and dark environments. *Environmental Science and Pollution Research*, **25**(21), 21224-21232. <https://doi.org/10.1007/s11356-018-2218-5>.
- Asada K. 2006. Production and scavenging of reactive oxygen species in chloroplasts and their functions. *Plant physiology*, **141**(2), 391-396. <https://doi.org/10.1104/pp.106.082040>.
- Asaeda T, and Imberger J. 1993. Structure of bubble plumes in linearly stratified environments. *Journal of Fluid Mechanics*, **249**, 35-57. <https://doi.org/10.1017/S0022112093001065>.

- Asaeda T, Rahman M, and Abeynayaka H D L. 2022a. "Hydrogen peroxide can be a plausible biomarker in cyanobacterial bloom treatment." *Scientific Reports* 12, no. 1: 1-11. | <https://doi.org/10.1038/s41598-021-02978-6>.
- Asaeda T, Rahman M, Liping X et al. 2022b. Hydrogen Peroxide Variation Patterns as Abiotic Stress Responses of *Egeria densa*. *Frontiers in Plant Science*.(2022), 4. <https://doi.org/10.3389/fpls.2022.855477>.
- Asaeda, T, Jayasanka S M D H, Xia L P et al. 2018. Application of hydrogen peroxide as an environmental stress indicator for vegetation management. *Engineering*, **4**(5), 610-616. <https://doi.org/10.1016/j.eng.2018.09.001>.
- Asaeda T, Senavirathna M D H J, Krishna, L. V et al. 2019. Impact of regulated water levels on willows (*Salix subfragilis*) at a flood-control dam, and the use of hydrogen peroxide as an indicator of environmental stress. *Ecological Engineering*, **127**, 96-102. <https://doi.org/10.1016/j.ecoleng.2018.10.028>.
- Asaeda T., Senavirathna M J, and Vamsi Krishna L. 2020. Evaluation of habitat preferences of invasive macrophyte *Egeria densa* in different channel slopes using hydrogen peroxide as an indicator. *Frontiers in plant science*, **11**, 422. <https://doi.org/10.3389/fpls.2020.00422>.
- Asada S, Fukuda K, Oh M, et al. 1999. Effect of hydrogen peroxide on the metabolism of articular chondrocytes. *Inflammation research*, **48**(7), 399-403. <https://doi.org/10.1007/s000110050478>.
- Athukorala D A, and Amarasinghe U S. 2010. Status of the fisheries in two reservoirs of the Walawe river basin, Sri Lanka: a case of participation of fishers in management. *Asian Fisheries Science*, **23**(3), 284-300. <https://doi.org/10.33997/j.afs.2010.23.3.002>
- Axler R P, and Owen C J. 1994. Measuring chlorophyll and phaeophytin: whom should you believe?. *Lake and Reservoir Management*, **8**(2), 143-151. <https://doi.org/10.1080/07438149409354466>.
- Babele P K, Kumar J, and Chaturvedi V. 2019. Proteomic de-regulation in cyanobacteria in response to abiotic stresses. *Frontiers in Microbiology*, **10**, 1315. <https://doi.org/10.3389/fmicb.2019.01315>.
- Barnuevo A, and Asaeda T. 2018. Integrating the ecophysiology and biochemical stress indicators into the paradigm of mangrove ecology and a rehabilitation blueprint. *PLoS One*, **13**(8), e0202227. <https://doi.org/10.1371/journal.pone.0202227>.
- Bouchard J N, and Purdie D A. 2011. Effect of elevated temperature, darkness and hydrogen peroxide treatment on oxidative stress and cell death in the bloom-forming toxic cyanobacterium *Microcystis aeruginosa*. *Journal of phycology*, **47**(6), 1316-1325. <https://doi.org/10.1111/j.1529-8817.2011.01074.x>.
- Cao H S, Kong F X, Luo L et al. 2006. Effects of wind and wind-induced waves on vertical phytoplankton distribution and surface blooms of *Microcystis aeruginosa* in Lake Taihu. *Journal of Freshwater Ecology*, **21**(2), 231-238. <https://doi.org/10.1080/02705060.2006.9664991>.
- Cao H S, Tao Y, Kong F X et al. 2008. Relationship between temperature and cyanobacterial recruitment from sediments in laboratory and field studies. *Journal of Freshwater Ecology*, **23**(3), 405-412. <https://doi.org/10.1080/02705060.2008.9664217>.
- Carmichael W W. 1981. Freshwater blue-green algae (cyanobacteria) toxins—A review. *The water environment*, 1-13. https://doi.org/10.1007/978-1-4613-3267-1_1.
- Caverzan A, Passaia G, Rosa S B et al. 2012. Plant responses to stresses: role of ascorbate peroxidase in the antioxidant protection. *Genetics and molecular biology*, **35**, 1011-1019. <https://doi.org/10.1590/S1415-47572012000600016>.
- Cheeseman J M. 2006. Hydrogen peroxide concentrations in leaves under natural conditions. *Journal of experimental botany*, **57**(10), 2435-2444. <https://doi.org/10.1093/jxb/erl004>.
- Chu Z, Jin X, Iwami N, and Inamori Y. 2007. The effect of temperature on growth characteristics and competitions of *Microcystis aeruginosa* and *Oscillatoria mougeotii* in a shallow, eutrophic lake simulator system. *Hydrobiologia*, **581**:217–223. <https://doi.org/10.1007/s10750-006-0506-4>.
- Cooper W J, and Lean D R. 1989. Hydrogen peroxide concentration in a northern lake: photochemical formation and diel variability. *Environmental science & technology*, **23**(11), 1425-1428. <https://doi.org/10.1021/es00069a017>.
- Dautania G K, and Singh G P. 2014. Role of light and dark cycle and different temperatures in the regulation of growth and protein expression in *Oscillatoria agardhii* strain. *Brazilian Archives of Biology and Technology*, **57**, 933-940. <https://doi.org/10.1590/S1516-8913201401970>.
- Derakhshan Z, Mahvi A H, Ehrampoush M H et al. 2018. Evaluation of kenaf fibers as moving bed biofilm carriers in algal membrane photobioreactor. *Ecotoxicology and environmental safety*, **152**, 1-7. <https://doi.org/10.1016/j.ecoenv.2018.01.024>.
- Dokulil M T, and Teubner K. 2000. Cyanobacterial dominance in lakes. *Hydrobiologia*, **438**(1), 1-12. <https://doi.org/10.1023/a:1004155810302>.

- Drake J L, Carpenter E J, Cousins M et al. 2010. Effects of light and nutrients on seasonal phytoplankton succession in a temperate eutrophic coastal lagoon. *Hydrobiologia*, **654**(1), 177-192. <https://doi.org/10.1007/s10750-010-0380-y>.
- Drábková M, Admiraal W, and Maršálek B. 2007a. Combined exposure to hydrogen peroxide and light selective effects on cyanobacteria, green algae, and diatoms. *Environmental science & technology*, **41**(1), 309-314. <https://doi.org/10.1021/es060746i>.
- Dumont S, and Rivoal J. 2019. Consequences of oxidative stress on plant glycolytic and respiratory metabolism. *Frontiers in Plant Science*, **10**, 166. <https://doi.org/10.3389/fpls.2019.00166>.
- Flombaum P, Gallegos J L, Gordillo R. A et al. 2013. Present and future global distributions of the marine Cyanobacteria Prochlorococcus and Synechococcus. *Proceedings of the National Academy of Sciences*, **110**(24), 9824-9829. <https://doi.org/10.1073/pnas.1307701110>.
- Foo S C, Chapman I J, Hartnell D M et al. 2020. Effects of H₂O₂ on growth, metabolic activity and membrane integrity in three strains of *Microcystis aeruginosa*. *Environmental Science and Pollution Research*, **27**(31), 38916-38927. <https://doi.org/10.1007/s11356-020-09729-6>.
- Gill S S, and Tuteja N. 2010. Reactive oxygen species and antioxidant machinery in abiotic stress tolerance in crop plants. *Plant physiology and biochemistry*, **48**(12), 909-930. <https://doi.org/10.1016/j.plaphy.2010.08.016>.
- Gorham P. R. 1964. Toxic algae as a public health hazard. *Journal-American Water Works Association*, **56**(11), 1481-1488. <https://doi.org/10.1002/j.1551-8833.1964.tb01355.x>.
- Grandgirard J., Poinot D., Krespi L et al. 2002. Costs of secondary parasitism in the facultative hyperparasitoid *Pachycrepoideus dubius*: does host size matter?. *Entomologia Experimentalis et Applicata*, **103**(3), 239-248. <https://doi.org/10.1046/j.1570-7458.2002.00982.x>.
- Gregor J, and Maršálek B. 2004. Freshwater phytoplankton quantification by chlorophyll a: a comparative study of in vitro, in vivo and in situ methods. *Water research*, **38**(3), 517-522. <https://doi.org/10.1016/j.watres.2003.10.033>.
- Harel Y, Ohad I, and Kaplan A. 2004. Activation of photosynthesis and resistance to photoinhibition in cyanobacteria within biological desert crust. *Plant Physiology* **136**(2), 3070-3079. <https://doi.org/10.1104/pp.104.047712>.
- Herrmann R. 1996. The daily changing pattern of hydrogen peroxide in New Zealand surface waters. *Environmental Toxicology and Chemistry: An International Journal*, **15**(5), 652-662. <https://doi.org/10.1002/etc.5620150507>.
- Imai H, Chang K H, Kusaba M et al. 2009. Temperature-dependent dominance of *Microcystis* (Cyanophyceae) species: *M. aeruginosa* and *M. wesenbergii*. *Journal of plankton research*, **31**(2), 171-178. <https://doi.org/10.1093/plankt/fbn110>.
- ISO 10260:1992. Water quality—measurement of biochemical parameters—spectrometric determination of the chlorophyll—a concentration. International Organization for Standardization, Geneva, Switzerland, 1992. <https://doi.org/10.3403/30303315>.
- Imteaz M A, and Asaeda T. 2000. Artificial mixing of lake water by bubble plume and effects of bubbling operations on algal bloom. *Water Research*, **34**(6), 1919-1929. [https://doi.org/10.1016/S0043-1354\(99\)00341-3](https://doi.org/10.1016/S0043-1354(99)00341-3).
- Joehnk K D, Huisman J E F, Sharples J et al. 2008. Summer heatwaves promote blooms of harmful cyanobacteria. *Global change biology*, **14**(3), 495-512. <https://doi.org/10.1111/j.1365-2486.2007.01510.x>.
- Jančula D, and Maršálek B. 2011. Critical review of actually available chemical compounds for prevention and management of cyanobacterial blooms. *Chemosphere*, **85**(9), 1415-1422. <https://doi.org/10.1016/j.chemosphere.2011.08.036>.
- Kaebernick M, Neilan B A, Börner T et al. 2000. Light and the transcriptional response of the microcystin biosynthesis gene cluster. *Applied and environmental microbiology*, **66**(8), 3387-3392. <https://doi.org/10.1128/AEM.66.8.3387-3392.2000>.
- Kura-Hotta M, Satoh K, and Katoh S. 1987. Relationship between photosynthesis and chlorophyll content during leaf senescence of rice seedlings. *Plant and Cell Physiology*, **28**(7), 1321-1329. <https://doi.org/10.1093/oxfordjournals.pcp.a077421>.
- Latifi A, Ruiz M, and Zhang C C. 2009. Oxidative stress in cyanobacteria. *FEMS microbiology reviews*, **33**(2), 258-278. <https://doi.org/10.1111/j.1574-6976.2008.00134.x>.
- Lea-Smith D J, Bombelli P, Vasudevan R et al. 2016. Photosynthetic, respiratory and extracellular electron transport pathways in cyanobacteria. *Biochimica et Biophysica Acta (BBA)-Bioenergetics*, **1857**(3), 247-255. <https://doi.org/10.1016/j.bbabi.2015.10.007>.

- LeBlanc Renaud S, Pick F R, and Fortin N. 2011. Effect of light intensity on the relative dominance of toxigenic and nontoxigenic strains of *Microcystis aeruginosa*. *Applied and environmental microbiology*, **77**(19), 7016-7022. <https://doi.org/10.1128/AEM.05246-11>.
- Leunert F, Eckert W, Paul A, Gerhardt V, & Grossart, H P. 2014. Phytoplankton response to UV-generated hydrogen peroxide from natural organic matter. *Journal of plankton research*, **36**(1), 185-197. <https://doi.org/10.1093/plankt/fbt096>.
- Liu M, Shi X, Chen C et al. 2017. Responses of *Microcystis* colonies of different sizes to hydrogen peroxide stress. *Toxins*, **9**(10), 306. <https://doi.org/10.3390/toxins9100306>.
- López C V G, García M D C C, Fernández F G A et al. 2010. Protein measurements of microalgal and cyanobacterial biomass. *Bioresource technology*, **101**(19), 7587-7591. <https://doi.org/10.1016/j.biortech.2010.04.077>.
- Machová K, Elster J, and Adamec L. 2008. Xanthophyceae assemblages during winter–spring flood: autecology and ecophysiology of *Tribonema fonticolum* and *T. monochloron*. *Hydrobiologia*, **600**(1), 155-168. <https://doi.org/10.1007/s10750-007-9228-5>.
- MacKintosh C, Beattie K A, Klumpp S et al. 1990. Cyanobacterial microcystin-LR is a potent and specific inhibitor of protein phosphatases 1 and 2A from both mammals and higher plants. *FEBS letters*, **264**(2), 187-192. [https://doi.org/10.1016/0014-5793\(90\)80245-E](https://doi.org/10.1016/0014-5793(90)80245-E).
- McLellan N L, and Manderville R A. 2017. Toxic mechanisms of microcystins in mammals. *Toxicology research*, **6**(4), 391-405. <https://doi.org/10.1039/c7tx00043j>.
- Ma J, and Wang P. 2021. Effects of rising atmospheric CO₂ levels on physiological response of cyanobacteria and cyanobacterial bloom development: A review. *Science of The Total Environment*, **754**, 141889. <https://doi.org/10.1016/j.scitotenv.2020.141889>.
- Ma Z, and Gao K. 2010. Spiral breakage and photoinhibition of *Arthrospira platensis* (Cyanophyta) caused by accumulation of reactive oxygen species under solar radiation. *Environmental and Experimental Botany*, **68**(2), 208-213. <https://doi.org/10.1016/j.envexpbot.2009.11.010>.
- Molle F, and Renwick M. 2005. *Economics and politics of water resources development: Uda Walawe irrigation project, Sri Lanka*, **87**. IWMI.
- Nishiyama Y, and Murata N. 2014. Revised scheme for the mechanism of photoinhibition and its application to enhance the abiotic stress tolerance of the photosynthetic machinery. *Applied microbiology and biotechnology*, **98**(21), 8777-8796. <https://doi.org/10.1007/s00253-014-6020-0>.
- Oliver N J, Rabinovitch-Deere C A, Carroll A L et al. 2016. Cyanobacterial metabolic engineering for biofuel and chemical production. *Current opinion in chemical biology*, **35**, 43-50. <https://doi.org/10.1016/j.cbpa.2016.08.023>.
- Oliveira C A, Machado S, Ribeiro R et al. 2014. Effect of light intensity on the production of pigments in *Nostoc* sp. *EJBMSR*, **2**(1), 23-36.
- Otsuka S, Suda S, Li R et al. 1999. Phylogenetic relationships between toxic and non-toxic strains of the genus *Microcystis* based on 16S to 23S internal transcribed spacer sequence. *FEMS Microbiology Letters*, **172**(1), 15-21. <https://doi.org/10.1111/j.1574-6968.1999.tb13443.x>.
- Otsuka S, Suda S, Li R et al. 2000. Morphological variability of colonies of *Microcystis* morphospecies in culture. *The Journal of general and applied microbiology*, **46**(1), 39-50. <https://doi.org/10.2323/jgam.46.39>.
- Ozawa K, Yokoyama A, Ishikawa K et al. 2003. Accumulation and depuration of microcystin produced by the cyanobacterium *Microcystis* in a freshwater snail. *Limnology*, **4**(3), 131-138. <https://doi.org/10.1007/s10201-003-0106-1>.
- Paerl H W, Xu H, McCarthy M J et al. 2011. Controlling harmful cyanobacterial blooms in a hyper-eutrophic lake (Lake Taihu, China): the need for a dual nutrient (N & P) management strategy. *Water research*, **45**(5), 1973-1983. <https://doi.org/10.1016/j.watres.2010.09.018>.
- Pham T L, and Utsumi M. 2018. An overview of the accumulation of microcystins in aquatic ecosystems. *Journal of environmental management*, **213**, 520-529. <https://doi.org/10.1016/j.jenvman.2018.01.077>.
- Preece E P, Hardy F J, Moore B C et al. 2017. A review of microcystin detections in estuarine and marine waters: environmental implications and human health risk. *Harmful Algae*, **61**, 31-45. <https://doi.org/10.1016/j.hal.2016.11.006>.
- Pojidaeva E, Zinchenko V, Shestakov S V et al. 2004. Involvement of the SppA1 peptidase in acclimation to saturating light intensities in *Synechocystis* sp. strain PCC 6803. *Journal of bacteriology*, **186**(12), 3991-3999. <https://doi.org/10.1128/JB.186.12.3991-3999.2004>.

- Queval G, Hager J, Gakiere B et al. 2008. Why are literature data for H₂O₂ contents so variable? A discussion of potential difficulties in the quantitative assay of leaf extracts. *Journal of experimental botany*, **59**(2), 135-146. <https://doi.org/10.1093/jxb/erm193>.
- Rajasekhar P, Fan L, Nguyen T et al. 2012. A review of the use of sonication to control cyanobacterial blooms. *Water research*, **46**(14), 4319-4329. <https://doi.org/10.1016/j.watres.2012.05.054>.
- Raja V, Majeed U, Kang H et al. 2017. Abiotic stress: Interplay between ROS, hormones and MAPKs. *Environmental and Experimental Botany*, **137**, 142-157. <https://doi.org/10.1016/j.envexpbot.2017.02.010>.
- Rantala A, Fewer D P, Hisbergues M et al. 2004. Phylogenetic evidence for the early evolution of microcystin synthesis. *Proceedings of the National Academy of Sciences*, **101**(2), 568-573. <https://doi.org/10.1073/pnas.030448910>.
- Raps S, Kycia J H, Ledbetter M C et al. 1985. Light intensity adaptation and phycobilisome composition of *Microcystis aeruginosa*. *Plant physiology*, **79**(4), 983-987. <https://doi.org/10.1104/pp.79.4.983>.
- Rastogi R P, Singh S P, Häder D P et al. 2010. Detection of reactive oxygen species (ROS) by the oxidant-sensing probe 2', 7'-dichlorodihydrofluorescein diacetate in the cyanobacterium *Anabaena variabilis* PCC 7937. *Biochemical and biophysical research communications*, **397**(3), 603-607. <https://doi.org/10.1016/j.bbrc.2010.06.006>.
- Rippka R, Deruelles J, Waterbury J B et al. 1979. Generic assignments, strain histories and properties of pure cultures of cyanobacteria. *Microbiology*, **111**(1), 1-61. <https://doi.org/10.1099/00221287-111-1-1>.
- Rahman, M., Asaeda, T., Fukahori, K., Imamura, F., Nohara, A., & Matsubayashi, M. (2023a). Hydrogen Peroxide Measurement Can Be Used to Monitor Plant Oxidative Stress Rapidly Using Modified Ferrous Oxidation Xylenol Orange and Titanium Sulfate Assay Correlation. *International Journal of Plant Biology*, **14**(3), 546-557. <https://doi.org/10.3390/ijpb14030043>.
- Rahman, M., Asaeda, T., Abeynayaka, H. D. L., & Fukahori, K. (2023b). An Assessment of the Effects of Light Intensities and Temperature Changes on Cyanobacteria's Oxidative Stress via the Use of Hydrogen Peroxide as an Indicator. *Water*, **15**(13), 2429. <https://doi.org/10.3390/w15132429>.
- Rosales-Loaiza N, Guevara M, Lodeiros C et al. 2008. Crecimiento y producción de metabolitos de la cianobacteria marina *Synechococcus* sp.(Chroococcales) en función de la irradiancia. *Revista de Biología Tropical*, **56**(2), 421-429. <https://doi.org/10.15517/rbt.v56i2.5596>.
- Saha R, Liu D, Hoynes-O'Connor A et al. 2016. Diurnal regulation of cellular processes in the cyanobacterium *Synechocystis* sp. strain PCC 6803: Insights from transcriptomic, fluxomic, and physiological analyses. *MBio*, **7**(3), e00464-16. <https://doi.org/10.1128/mBio.00464-16>.
- Salonen K, Sarvala J, Järvinen M et al. 1999. Phytoplankton in Lake Tanganyika—vertical and horizontal distribution of in vivo fluorescence. In *From Limnology to Fisheries: Lake Tanganyika and Other Large Lakes* (pp. 89-103). Springer, Dordrecht. https://doi.org/10.1007/978-94-017-1622-2_9
- Satterfield C N, and Bonnell A H. 1955. Interferences in titanium sulfate method for hydrogen peroxide. *Analytical chemistry*, **27**(7), 1174-1175. <https://doi.org/10.1021/ac60103a042>.
- Schwimmer M. 1968. Medical aspects of phycology. *Algae, man and environment*, 279-358.
- Scully N M, McQueen D J, and Lean D R S. 1996. Hydrogen peroxide formation: the interaction of ultraviolet radiation and dissolved organic carbon in lake waters along a 43–75 N gradient. *Limnology and Oceanography*, **41**(3), 540-548. <https://doi.org/10.4319/lo.1996.41.3.0540>.
- Sharma P, Jha A B, Dubey R S et al. 2012. Reactive oxygen species, oxidative damage, and antioxidative defense mechanism in plants under stressful conditions. *Journal of botany*, 2012. <https://doi.org/10.1155/2012/217037>.
- Taras M J. 1971. Standard methods for the examination of water and wastewater. American Public Health Association: New York, 13th edition. 874 P.
- Ugya A Y, Imam T S, Li A et al. 2020. Antioxidant response mechanism of freshwater microalgae species to reactive oxygen species production: a mini review. *Chemistry and Ecology*, **36**(2), 174-193. <https://doi.org/10.1080/02757540.2019.1688308>.
- Utkilen H, and Gjørlme N. 1992. Toxin production by *Microcystis aeruginosa* as a function of light in continuous cultures and its ecological significance. *Applied and environmental microbiology*, **58**(4), 1321-1325. <https://doi.org/10.1128/aem.58.4.1321-1325.1992>.
- Vestervik P S, Misiorek J O, Spoof L E et al. 2012. Comparative cellular toxicity of hydrophilic and hydrophobic microcystins on Caco-2 cells. *Toxins*, **4**(11), 1008-1023. <https://doi.org/10.3390/toxins4111008>.
- Visser P M, Ibelings B W, Bormans M et al. 2016. Artificial mixing to control cyanobacterial blooms: a review. *Aquatic Ecology*, **50**(3), 423-441. <https://doi.org/10.1007/s10452-015-9537-0>.

- Walsh K, Jones G J, and Dunstan R H. 1997. Effect of irradiance on fatty acid, carotenoid, total protein composition and growth of *Microcystis aeruginosa*. *Phytochemistry*, **44**(5), 817-824. [https://doi.org/10.1016/S0031-9422\(96\)00573-0](https://doi.org/10.1016/S0031-9422(96)00573-0).
- Weerakoon H P A T, Atapaththu K S S, and Asanthi H B. 2018. Toxicity evaluation and environmental risk assessment of 2-methyl-4-chlorophenoxy acetic acid (MCPA) on non-target aquatic macrophyte *Hydrilla verticillata*. *Environmental Science and Pollution Research*, **25**(30), 30463-30474. <https://doi.org/10.1007/s11356-018-3013-z>.
- Welkie D G, Rubin B E, Diamond S et al. 2019. A hard day's night: cyanobacteria in diel cycles. *Trends in microbiology*, **27**(3), 231-242. <https://doi.org/10.1016/j.tim.2018.11.002>.
- Wu Z, Song L, and Li R. 2008. Different tolerances and responses to low temperature and darkness between waterbloom forming cyanobacterium *Microcystis* and a green alga *Scenedesmus*. *Hydrobiologia*, **596**(1), 47-55. <https://doi.org/10.1007/s10750-007-9056-7>.
- Yoshida M, Yoshida T, Satomi M et al. 2008. Intra-specific phenotypic and genotypic variation in toxic cyanobacterial *Microcystis* strains. *Journal of applied microbiology*, **105**(2), 407-415. <https://doi.org/10.1111/j.1365-2672.2008.03754.x>.
- Yoshizawa S, Matsushima R, Watanabe M F et al. 1990. Inhibition of protein phosphatases by microcystis and nodularin associated with hepatotoxicity. *Journal of cancer research and clinical oncology*, **116**(6), 609-614. <https://doi.org/10.1007/bf01637082>.
- You J, Mallery K, Hong J et al. 2018. Temperature effects on growth and buoyancy of *Microcystis aeruginosa*. *Journal of Plankton Research*, **40**(1), 16-28. <https://doi.org/10.1093/plankt/fbx059>.
- Zhou B, Wang J, Guo Z et al. 2006. A simple colorimetric method for determination of hydrogen peroxide in plant tissues. *Plant Growth Regulation*, **49**(2), 113-118. <https://doi.org/10.1007/s10725-006-9000-2>.